

Mapping supply of and demand for ecosystem services to assess environmental justice in New York City

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Abstract. Livability, resilience, and justice in cities are challenged by climate change and the historical legacies that together create disproportionate impacts on human communities. Urban green infrastructure has emerged as an important tool for climate change adaptation and resilience given their capacity to provide ecosystem services such as local temperature regulation, stormwater mitigation, and air purification. However, realizing the benefits of ecosystem services for climate adaptation depend on where they are locally supplied. Few studies have examined the potential spatial mismatches in supply and demand of urban ecosystem services, and even fewer have examined supply–demand mismatches as a potential environmental justice issue, such as when supply–demand mismatches disproportionately overlap with certain socio-demographic groups. We spatially analyzed demand for ecosystem services relevant for climate change adaptation and combined results with recent analysis of the supply of ecosystem services in New York City (NYC). By quantifying the relative mismatch between supply and demand of ecosystem services across the city we were able to identify spatial hot- and coldspots of supply–demand mismatch. Hotspots are spatial clusters of census blocks with a higher mismatch and coldspots are clusters with lower mismatch values than their surrounding blocks. The distribution of mismatch hot- and coldspots was then compared to the spatial distribution of socio-demographic groups. Results reveal distributional environmental injustice of access to the climate-regulating benefits of ecosystem services provided by urban green infrastructure in NYC. Analyses show that areas with lower supply–demand mismatch tend to be populated by a larger proportion of white residents with higher median incomes, and areas with high mismatch values have lower incomes and a higher proportion of people of color. We suggest that urban policy and planning should ensure that investments in “nature-based” solutions such as through urban green infrastructure for climate change adaptation do not reinforce or exacerbate potentially existing environmental injustices.

Key words: cities; climate change adaptation; regulating ecosystem services; resilience; spatial analysis; urban ecosystem services.

INTRODUCTION

Climate change, urban green infrastructure, and ecosystem services

Climate change is already impacting cities and both current and future risks affect communities unequally (Pelling and Garschagen 2019). Because of the concentration of people and infrastructure in urban areas (Bouwer 2010, Dickson et al. 2012, Depietri and McPhearson 2017, Depietri et al. 2018), cities face disproportionate current and future risks from increased heat, heat waves, and more frequent and intense flooding by extreme

weather events. In addition, climate change is expected to have detrimental effects on air quality (Ebi and McGregor 2008, Kinney 2008, Jacob and Winner 2009), worsening the impact of air pollution on human health (Rosenzweig et al. 2010, Revi et al. 2014). Urban green infrastructure (UGI) is increasingly being implemented as an alternative to traditional engineered approaches for improving urban resilience to climate change impacts (Nilon et al. 2017). In this paper, we take a broad definition of UGI as the network of planned and unplanned green spaces that provide ecosystem services through the support of ecological functions, also described as a type of urban ecological infrastructure (Childers et al. 2019). By focusing on the strategic role of green spaces and ecosystem services in urban planning, UGI falls within the umbrella term of nature-based solutions (NBS) (Kabisch et al. 2017), defined as “actions to protect, sustainably manage and restore natural or modified

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ecosystems, which address societal challenges (e.g., climate change, food and water security, or natural disasters) effectively and adaptively, while simultaneously providing human well-being and biodiversity benefits” (Cohen-Shacham et al. 2016:xii). As a pioneer city embracing UGI as a promising climate adaptation tool, New York City (NYC) has built green roofs, installed thousands of bioswales, planted a million trees, and invested in other forms of UGI in the last decade to combat water quality, urban flooding, heat, and air-quality challenges (Campbell et al. 2014, McPhearson et al. 2013a, New York City 2017, 2019). Many other cities are doing the same from Philadelphia, Pennsylvania to Portland, Oregon, to Chinese cities embracing the “sponge” cities concept to deploy UGI as a stormwater solution (Wang et al. 2018, City of Philadelphia n.d., The City of Portland n.d.).

The ecosystem services (ES) conceptual framework has been widely used to articulate the climate regulatory benefits of UGI in urban climate adaptation and resilience planning (Demuzere et al. 2014, Hansen and Pauleit 2014, Hansen et al. 2015, McPhearson et al. 2015, Geneletti et al. 2020). ES are broadly defined as “the direct and indirect contributions of ecosystems to human well-being” (Groot et al. 2012), and are commonly classified in categories such as provisioning, regulating, cultural, and habitat ES (Millennium Ecosystem Assessment 2003, The Economics of Ecosystems and Biodiversity (TEEB) 2008, Gómez-Baggethun et al. 2013). Regulating ES, including local temperature regulation, runoff mitigation, and air purification are three of the most important services in UGI planning for urban climate resilience and adaptation (Hansen et al. 2015, Kabisch et al. 2017).

UGI and ES are spatially explicit. UGI studies focusing on parks (Rigolon 2016, Rigolon et al. 2018) or vegetation (Nesbitt et al. 2019) have shown that the distribution of green assets across the city is uneven. Consequently, the (uneven) distribution of UGI affects the supply of ES. Different ES require different spatial relations between the service-providing areas and their targeted beneficiaries (Fisher et al. 2009, Burkhard et al. 2014, Andersson et al. 2015). For example, most ES under the provisioning category (e.g. food or timber production) may have a decoupled spatial relation between the service providing area and their beneficiaries, given that the service can be transported. However, some ES, such as the three regulating services considered key for urban resilience (local temperature regulation, stormwater runoff mitigation, and air purification), cannot be actively transported. In these cases, ES providing and benefiting areas need to overlap spatially for their benefits to be delivered. Thus, local demand for regulating ES, including local temperature regulation, stormwater runoff mitigation, and air purification needs to be met at the local level (Burkhard et al. 2014, Hamstead et al. 2016). Recognizing potential mismatches between where ES are supplied and where demand is higher is essential

when planning and creating policy for UGI investments to ensure that benefits are provided where they are most needed so that investments can maximize impact (Burkhard et al. 2012, McPhearson et al. 2013b, Keeler et al. 2019a).

Although the mapping of ES supply is more developed in ES research, mapping ES demand is a relatively new concept that tends to be overlooked or taken for granted (Burkhard et al. 2014, Wolff et al. 2015, Cortinovis and Geneletti 2018, Keeler et al. 2019a). In NYC, for example, the supply of ES has been mapped by Kremer et al. (2016), but the spatial dynamics between supply and demand have yet to be considered. Assessing the supply of regulating ES is challenged by a lack of observational data, leading ES researchers to rely on process-based modeling to estimate the potential benefits that ecosystems could provide (Wolff et al. 2015). In this paper, we follow (Burkhard and Maes 2017:185) and define ES supply as “the capacity of ecosystems to provide ecosystem services.” The demand of regulating ES is usually conceptualized as “need for risk reduction” (Wolff et al. 2015), and, with a similar dearth of observational data, tends to rely on proxy indicators in assessments.

Environmental justice of ecosystem services distribution

Considering supply as the potential benefits that ecosystems could provide, and demand as the need for these benefits based on the urge to alleviate environmental risks, the distribution of mismatches between supply and demand is an important step in revealing potential distributional environmental injustices. Certain socio-demographic groups are, through historical racism and other legacies of environmental and social injustice, more prone to having their needs for regulating ES unsatisfied (Kabisch and Haase 2014, Rigolon 2016, Rigolon et al. 2018). Distributional environmental justice has been broadly defined as “the spatial distribution of environmental goods and ills amongst people” (Ernstson 2013:8). Initially focused on the allocation of toxic activities such as dumpsites and industrial facilities close to low-income communities and communities of color (Bullard 2008, Brender et al. 2011), environmental justice has evolved as a body of research and advocacy to incorporate the exposure to environmental hazards such as flooding or extreme heat (Maantay and Maroko 2009, Jenerette et al. 2011, Collins et al. 2018, Herreros-Cantis et al. 2020) and the unequal investments in beneficial interventions such as UGI and open spaces (Miyake et al. 2010, Kabisch and Haase 2014, Rigolon 2016, Rigolon et al. 2018). Distributional environmental justice in the United States is intimately linked to the planning and housing policies undertaken by public and private institutions during the 20th century and that caused communities of color, especially African American, to remain segregated, underserved, and discriminated against (Rothstein 2017, Nelson et al. n.d.). A

prime example of these practices is known as red lining, a process by which banks systematically neglected to provide loans and mortgages in neighborhoods based on racial composition. While the mortgages were neglected by banks, the drawing of the red-lining maps was performed by the Home Owners' Loan Corporation, a federally backed institution. The effects of red lining have been linked to current distributional injustices in several American cities and are spatially correlated with the distribution of green spaces and environmental hazards in communities of color (Grove et al. 2018, Hoffman et al. 2020).

Assessments on UGI's distributional justice often compare differences in distance to green areas, green surface proportion, or recreational quality of parks between different socio-demographic groups (Rigolon 2016). However, the links between distributional environmental justice and ES remains poorly studied. We suggest that linking these two areas of study adds an important dimension of ecological functioning to simpler examinations of the spatial distribution of UGI, such as parks or urban greenery. For example, Graça et al. (2017) analyzed the relationship between several ES supply indicators and socio-demographic factors including age distribution, education level, building's age and tenancy. However, this study did not consider ES demand and its effect on the relevance of the distribution of ES supply. Baró et al. (2019) analyzed the distributional justice of regulating ES supply in Barcelona through the i-tree model, also without considering ES demand. A supply-demand assessment was carried out for the same city in Baró et al. (2016), but environmental justice was not considered in this prior study. These examples are not intended to lessen the impact of such research, rather to note the need to address the supply and demand perspective as well as the environmental justice perspective together as two important and consistently missing dimensions in urban ecosystem services research. In contrast to the regulating services we examine here, several studies have focused on the justice aspects of cultural ES (Amorim Maia et al. 2020, Łaskiewicz and Sikorska 2020, Suárez et al. 2020).

Here, we bring environmental justice dimensions and the differential need for ES together in the same study to understand not only how ES demand compares with supply across space, but also how socio-demographic indicators used in distributional justice studies can bring justice perspectives and concerns more fully into ES research and practice. Understanding whether the distribution of UGI and their ES benefit communities most in need is a key starting point for improving UGI planning to address issues of social inequity and environmental injustice (Marshall and Gonzalez-Meler 2016). Depriving neighborhoods from regulating ES may lead to further perpetuation of historical inequalities (Reckien et al. 2017), especially as the impacts of climate change in cities increase risk of flooding or heat waves in vulnerable communities and thus create new

environmental justice challenges (Depietri et al. 2018, Pelling and Garschagen 2019).

Research objectives

We use New York City (NYC) as a case study where both environmental justice and ecosystem services research has been well explored, but to date poorly explicitly linked. Additionally, data availability, recent investments in UGI, ongoing climate impacts, and historical environmental injustices make NYC a useful empirical case to investigate the environmental justice implications of potential ES supply-demand mismatches. Here we conceptually and empirically link urban ES supply and demand with questions of distributional environmental justice.

This study has two main objectives: First, we generate ES demand maps for each of the regulation ES considered key for climate change adaptation and resilience, including local temperature regulation, runoff mitigation, and air purification. In addition, we follow previously published methods (Kremer et al. 2016) to map ES supply, with minor adjustments and improvements. Second, we analyze the distributional justice of current ES in NYC by comparing the distribution of supply-demand mismatch hotspots with the distribution of different socio-demographic groups across the city. Combining these spatial analytical approaches, we provide a quantitative assessment of distributional injustices with respect to regulating ES provided by UGI in NYC.

The ES studied in this research are a subset of the many that multifunctional UGI may provide (Gómez-Baggethun et al. 2013, Keeler et al. 2019b). However, we focus on temperature regulation, runoff mitigation, and air quality regulation because these are common goals for city investments in UGI for climate change risk reduction (Kabisch et al. 2017) including in NYC (Depietri et al. 2018).

METHODS

Study area

With over eight million residents and a population density that reaches over 10,000 people/km², NYC is the largest and most dense city in the United States. This makes NYC an epicenter for examining climate change impacts given studies that suggest environmental challenges are expected to worsen in the context of climate change (González et al. 2019). For example, projections indicate that NYC's mean temperatures could rise by as much as 7.5°F by 2080 (Horton et al. 2010).

NYC is composed of five different boroughs (Manhattan, Queens, Bronx, Brooklyn, and Staten Island) and 59 community districts. Community districts (CDs) are a key planning unit for multiple city government agencies, because local decision making is usually carried out at this administrative scale (Kremer et al. 2016, New York

City’s Mayor Community Affairs Unit (NYCMCAU) n.d.). In this study, we analyze supply–demand mismatches at the census block level, which is the smallest spatial unit with population data available. The City of New York has a total of 38,768 census blocks, with a mean area of 2.05 ha. Additionally, 30,131 of the city’s census blocks are inhabited, with a mean population of 271 people and a mean population density of 176 people per hectare.

A formerly redlined city (Fig. 1), NYC has a legacy of environmental injustice that has been studied both

quantitatively and qualitatively. For example, Miyake et al. (2010) found that factors such as recreational quality and park acreage showed a significant relationship with socio-demographic factors, concluding that people of color had access to only smaller and lower-quality parks. Other studies have focused on the responses, effects, and drivers of the zoning of noxious land uses and activities that drive environmental injustice in the city (Brown et al. 2003, Sze 2006).

In terms of green space, NYC has a high density of natural land cover, but it is unequally distributed in

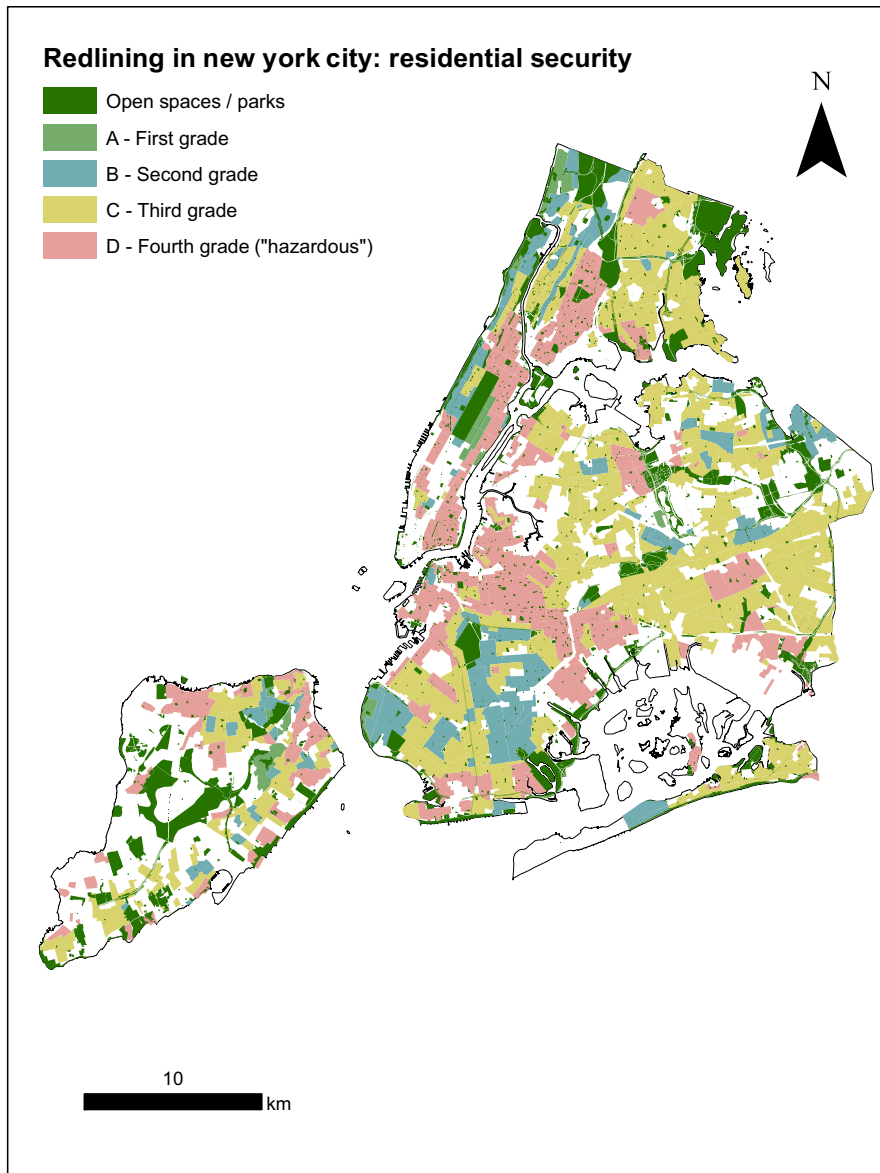


FIG. 1. Red lining map of New York City created by the Home Owners’ Loan Corporation during the 1930s. Residential neighborhoods were given a mortgage security grade that reflected the security of a potential investment made by banks and other mortgage lenders. While grade A refers to low risk areas, grade D refers to areas qualified as “hazardously” risky. Data and description obtained from (Nelson et al. n.d.).

TABLE 1. Percentage of each land cover category per borough in New York City according to the land cover classification developed in 2010 by MacFaden et al. (2012).

	Manhattan	Queens	Bronx	Brooklyn	Staten Island
Tree canopy	19	18	23	16	29
Grass/shrub	7	19	16	13	27
Bare earth	0	2	1	1	2
Water body	1	1	1	1	2
Building	32	19	19	26	11
Road	18	17	16	15	11
Other (impervious)	21	24	24	28	19

TABLE 2. Proportion of each race/ethnicity per borough in New York City according to the decennial census 2010 (U.S. Census Bureau *n.d.* a, b).

Race/ethnicity	Manhattan	Queens	Bronx	Brooklyn	Staten Island
White	0.480	0.276	0.109	0.357	0.640
Hispanic/Latino	0.254	0.275	0.535	0.198	0.173
Black/African American	0.129	0.177	0.301	0.319	0.095
American Indian/Native Alaskan	0.001	0.003	0.002	0.002	0.001
Asian	0.112	0.228	0.034	0.104	0.074
Native Hawaiian/Pacific Islander	0.000	0.000	0.000	0.000	0.000
Other race	0.003	0.014	0.006	0.004	0.002
Two or more races	0.019	0.025	0.012	0.016	0.014

space at the borough level (Table 1). The distribution of different races and ethnicities is also uneven across the five boroughs of the city (Table 2). Figs. 2, 3 show the distribution of race and income at the census block level. The overlap of race and income at this finer resolution with the distribution of green spaces demonstrates the uneven distribution of green space across socio-demographic groups.

Despite the importance given to UGI and the city's role as a global leader in the inclusion of ES thinking in its planning policies (Hansen et al. 2015), the environmental justice implications of potential supply–demand mismatches for ES have not been examined at a citywide, continuous scale. Additionally, though the supply of specific ES has been studied in NYC (Kremer et al. 2016), the role of ES demand has not been explored. This NYC case study aims to develop an empirical approach based on previously published conceptual papers that can help prioritize investments in UGI for climate adaptation in areas where they are most needed. We provide a methodological template for use in other cities facing similar challenges and considering or implementing similar UGI solutions.

Mapping ES demand

Our goal in mapping ES demand is to identify the areas in the city with the highest need for each regulating ES (Burkhard et al. 2014). We define demand for ES as “need for risk reduction” following Wolff et al. (2015)

and follow methods elaborated by Baró et al. (2016) to map ES demand, as they did in Barcelona. Methods consist of developing a cross-tabulation matrix that combines two risk factors to generate a demand index that ranges from 0 (not relevant demand) to 1 (very highly relevant demand). The factors considered for each ES are population density per census block (exposure) and a service-specific hazard factor. As in Baró et al. (2016), we assume the role of population density in the exposure to hazards as constant.

For local temperature regulation, the service-specific hazard factor for the cross-tabulation matrix was land surface temperature (Table 3). Break values for demand within the matrix were defined using the heat index thresholds developed by the National Weather Service and referenced in “NYC’s Risk Landscape: A Guide to Hazard Mitigation” available in the NYC Emergency Management portal (National Weather Service (NWS) *n.d.*, New York City Emergency Management (NYCEM) *n.d.*). The heat index provides thresholds of different degrees of risk due to exposure to heat based on apparent temperature, which combines temperature and relative humidity (Table 4). Relative humidity across NYC was assumed as 70% after calculating the average relative humidity for the months of June, July, and August during the years 1987–2017 for the meteorological data retrieved from the NOAA NYC weather stations located in Central Park (70%), Kennedy airport (71%), and LaGuardia airport (65%) (National Oceanic and Atmospheric Administration (NOAA) *n.d.*). Land

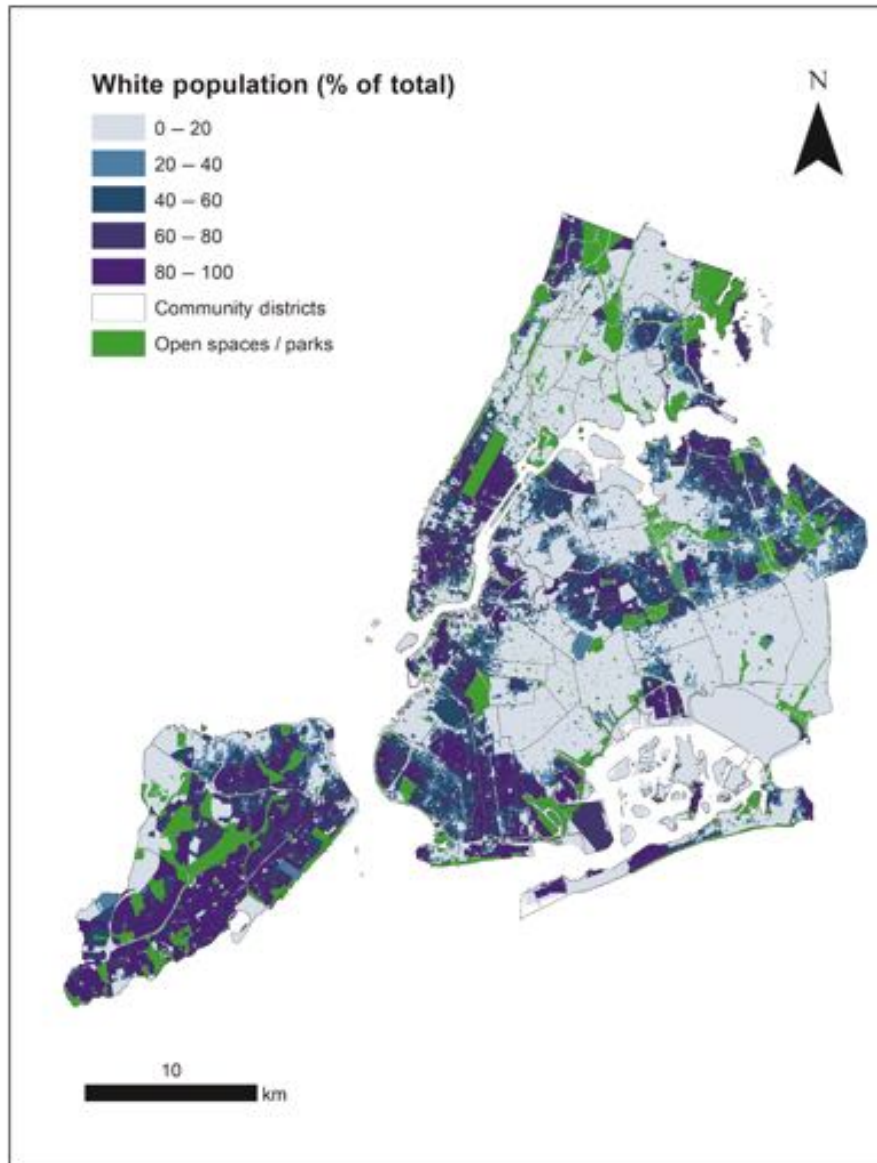


FIG. 2. Ratio of total population classified as white according to the 2010 decennial census, per census block (U.S. Census Bureau n.d. *a, b*). Additionally, the distribution of parks in New York City (Department of Information Technology & Telecommunications (DoITT) n.d.) is included to depict the spatial correlation between race and public green space visually.

surface temperature was obtained from Landsat 7's band 6_1 (low gain thermal band, sampled on a 60×60 m resolution). As in Imhoff et al. (2010), a series of scenes were compiled to generate an "average summer day" in NYC. The scenes ($n = 13$) were selected considering the year (2008–2012, in order to be consistent with the land-cover cartography used in the study, developed in 2010), month (June, July, and August) and cloud cover (only scenes with cloud cover lower than 10% were considered). Table 5 shows the dates and extreme values recorded in each scene. Besides creating an "average summer day," this methodology allowed for

filling the gaps caused by the sensor failure of Landsat 7 through the calculation of mean values per cell.

For runoff mitigation, we mapped demand considering the percentage of impervious surfaces per census block as a hazard factor (Table 6). We chose this because of the lack of accessible data and resources to develop a reliable modeling approach of the simulate runoff in NYC (Rosenzweig et al. 2020). Impervious surface is known to impact the water cycle by increasing the amount and speed of runoff generated (Shuster et al. 2005). In addition, this indicator has been previously used to assess demand for flood protection (Liquete

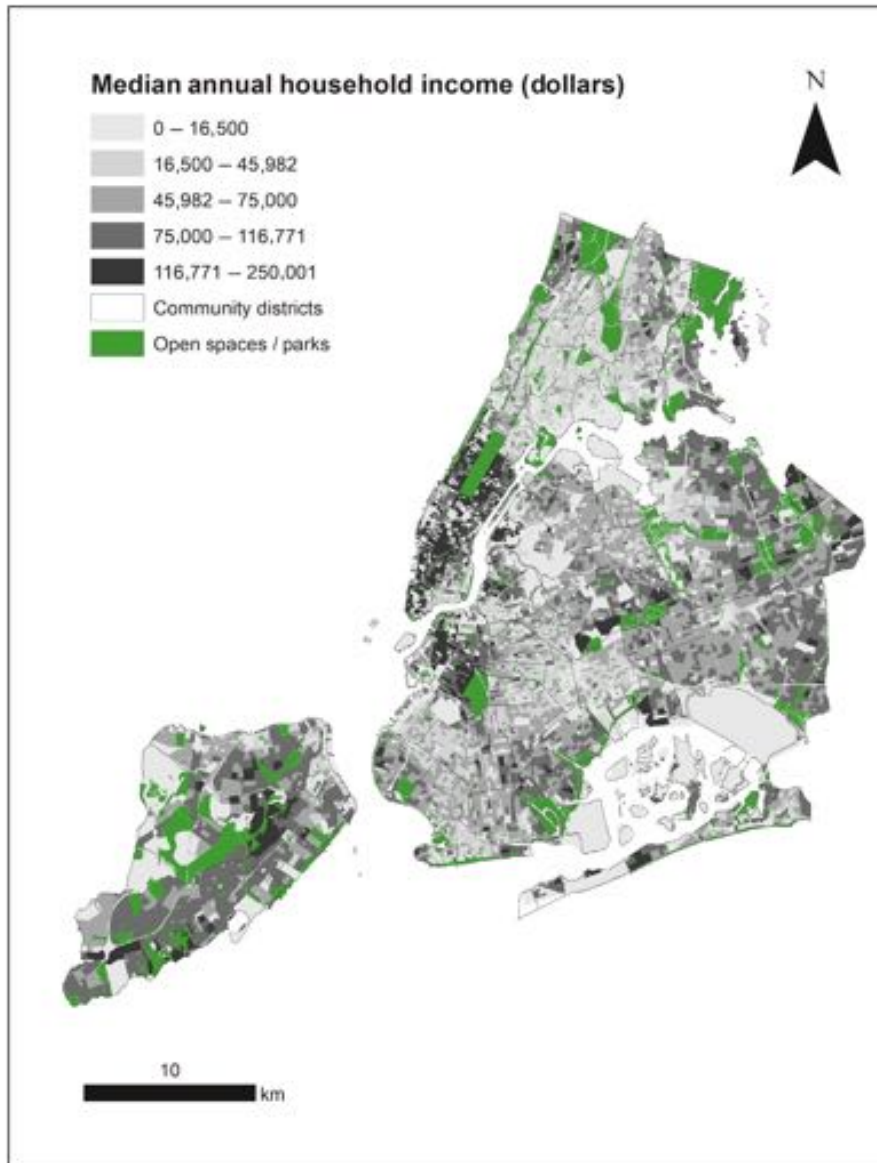


Fig. 3. Median income per census block according to the American Community Survey 5-yr estimates 2013–2017 (U.S. Census Bureau n.d. *a, b*), with the distribution of public green (Department of Information Technology & Telecommunications (DoITT) *n.d.*) spaces overlapped. Values range from 0 (no income) to 1.0 (maximum estimated income in New York City).

et al. 2013). An equal interval approach is taken to define the break values, setting maximum demand when impervious surface exceeds 80%. The information on the proportion of impervious surface per census block was gathered by using the land cover map developed by MacFaden et al. (2012).

To define demand for air purification, data on predicted average concentrations of NO_2 and O_3 in 2010 were used to estimate the air pollution hazard. These data were retrieved from the data set New York City Community Air Survey (NYCCAS) Air Pollution Rasters (Department of Health and Mental Hygiene 2017).

NO_2 and O_3 were chosen because of data availability to calculate ES supply, but future iterations of this project may incorporate other pollutants, such as particulate matter, if possible. We assessed each pollutant separately rather than combining both compounds in one single service for two reasons. To begin with, the data available for these pollutants were not temporally consistent. For NO_2 , the data provide yearly average concentrations, and for O_3 only summer concentrations are provided. In addition, the pollutants considered have complex dynamics that determine their occurrence. For example, O_3 is a secondary pollutant that results from the

TABLE 3. Cross-tabulation matrix used to determine the demand for the ecosystem service “local temperature regulation” at the census block level.

Population density (inhabitants/ha)	Temperature (Fahrenheit)					
	<75	75–80	80–85	85–95	95–100	>100
<5	0	0	0	0	0	0
5–50	0	0.2	0.2	0.4	0.4	0.6
50–100	0	0.2	0.4	0.4	0.6	0.8
100–200	0	0.4	0.4	0.6	0.6	0.8
200–400	0	0.4	0.6	0.6	0.8	0.8
>400	0	0.6	0.8	0.8	0.8	1

Notes: Population density and temperature are used as exposure and hazard indicators to define the need for mitigating the risks associated with extreme heat. The break values for population density are extracted from Baró et al. (2016), while the break values for temperature are based on the thresholds defined by the heat index used by New York City Emergency Management (NYCEM *n.d.*). The temperature breaks in Celsius are 23.89°C (75°F), 26.67°C (80°F), 29.44°C (85°F), 35.00°C (95°F), and 37.78°C (100°F).

interaction between NOx, VOCs, and specific meteorological conditions (Pun et al. 2003). Because of this, conflating both pollutants as if they both occurred at the same time scale is not appropriate. Data were resampled from their original resolution (300 × 300 m) to 1 m using a bilinear interpolation method to generate mean concentration per census block. Break values were defined considering the maximum tolerable concentrations allowed by the National Ambient Air Quality Standards (U.S. Environmental Protection Agency (EPA) 2014). These quality standards (53 ppb for mean annual NO₂ and 70 ppb for 8-h O₃) were used in the matrix to set the break value for the highest demand index, and then we equally subdivided this number to define the lower demand break points (Tables 7, 8).

Mapping ES supply

To map ES supply, we followed the methods originally developed in Kremer et al. (2016). In this study, the supply of ES in NYC was mapped by relying on a raster-based approach that combined a high-resolution (1 × 1 m) land cover map (MacFaden et al. 2012) with other sources of secondary data to create supply indicators for each ES (Table 9). The methods and data used

are presented in Appendix S1: Section S1 and mimic the procedure presented in Kremer et al. (2016). A minor adjustment was done to the methodology for mapping the supply of the local temperature regulation ES. Instead of using one single Landsat scene to calculate the reduced temperature due to the natural land cover, the “average summer day” data generated in the demand assessment was used to avoid relying on a single temperature record.

We then generated a series of 1 × 1 m raster maps showing a normalized supply value ranging from 0 (no supply at all) to 1 (maximum supply). In order to compare ES supply with demand, it is important that supply data is aggregated to the census block level. An average supply value was calculated by considering the area within each census block and an additional 400 m service area generated through the Network Analyst extension available in ArcGIS 10 (ESRI, Redlands, CA, USA). To generate service areas for each census block, we built a network data set with the walkable roads of the data set “NYC Street Centerline” (Department of Information Technology and Telecommunications 2014). A 400-m service area incorporates the idea that the residents of a given census block may be able to access ecosystem services supplied outside of their block and benefit from them. For example, residents are often exposed to urban flooding, air quality, or heat hazards while walking or biking to a supermarket or to work or school. We chose a 400-m area as a conservative approach, because this is the lowest distance normally considered in walkability assessments (Miyake et al. 2010).

Comparing supply and demand: the spatial supply–demand mismatch

Each census block in NYC was assigned a value for supply and for demand that ranged from 0 to 1. For supply, the value represents the potential benefits provided by ecosystems on a normalized scale. For demand, the value indicates relevant need for risk reduction associated with the specific ES mapped.

To assess the mismatch between supply and demand across the city, we generated a supply–demand mismatch value per census block. In Burkhard et al. (2012), a supply–demand subtraction is suggested to represent the “budget” of each ES per land cover. This subtraction

TABLE 4. Temperature intervals for different likelihoods of heat disorders due to prolonged exposure or strenuous activity under relative humidity conditions of 70% (NWS *n.d.*, New York City Emergency Management *n.d.*).

Temperature interval	80–85°F	85–95°F	95–100°F	>100°F
Heat stress risk	Caution—possible fatigue with prolonged exposure and/or physical activity	Extreme caution—sunstroke, heat cramps and heat exhaustion possible with prolonged exposure and/or physical activity	Danger—sunstroke, heat cramps and heat exhaustion likely, and heatstroke possible with prolonged exposure and/or physical activity	Extreme danger—heat/sunstroke highly likely with continued exposure

TABLE 5. Summary of the Landsat 7 scenes used to elaborate an average summer land surface temperature map.

Date of scene	Minimum temperature value in raster (°C)	Maximum temperature value in raster (°C)
29 July 2008	10.0	46.7
7 June 2009	12.9	47.6
25 July 2009	10.6	43.7
10 August 2009	17.2	40.6
26 August 2009	12.3	42.4
3 July 2010	15.1	55.0
20 August 2010	12.9	43.7
29 August 2010	10.0	52.2
15 July 2011	0.5	50.5
22 July 2011	16.7	41.9
23 August 2011	-0.2	48.4
15 June 2012	13.4	48.4
1 July 2012	17.2	45.9

TABLE 6. Cross-tabulation matrix used to determine the demand for the ecosystem service “runoff mitigation” at the census block level.

Population density (inhabitants/ha)	Impervious surface (%)					
	<10	10–20	20–40	40–60	60–80	>80
<5	0	0	0	0	0	0
5–50	0	0.2	0.2	0.4	0.4	0.6
50–100	0	0.2	0.4	0.4	0.6	0.8
100–200	0	0.4	0.4	0.6	0.6	0.8
200–400	0	0.4	0.6	0.6	0.8	0.8
>400	0	0.6	0.8	0.8	0.8	1

Notes: Population density and % of the block’s surfaces being impervious are used as exposure and hazard indicators to define the need for mitigating the risks associated with urban flooding. The break values for population density are extracted from Baró et al. (2016), and the break values for impervious surfaces were set using an equal interval (20%) from 0 to 100%.

serves as a proxy for the deviation between the ES provided and the relevance of their need. Consequently, mismatch was calculated in this paper by subtracting the supply index from the demand index. Results ranged from 1 to -1, with 1 indicating the highest (negative) mismatch (maximum possible demand and absence of supply). That is, higher values represent areas in which the demand reflects a more relevant need for ES, but supply is low in comparison with other parts of the study area.

Comparing supply–demand mismatch and socio-demographic groups

In the final step we compared the distribution of mismatch values to that of two socio-demographic indicators, including (1) the percentage of different races and ethnicities, and (2) the normalized median annual household income. When comparing the distribution of

TABLE 7. Cross-tabulation matrix used to determine the demand for the ecosystem service “Air purification (NO₂)” at the census block level.

Population density (inhabitants/ha)	NO ₂ concentration (ppb)					
	<9	9–18	18–25	25–36	36–53	>53
<5	0	0	0	0	0	0
5–50	0	0.2	0.2	0.4	0.4	0.6
50–100	0	0.2	0.4	0.4	0.6	0.8
100–200	0	0.4	0.4	0.6	0.6	0.8
200–400	0	0.4	0.6	0.6	0.8	0.8
>400	0	0.6	0.8	0.8	0.8	1

Notes: Population density and mean annual NO₂ concentration are used as exposure and hazard indicators to define the need for mitigating the risks associated with NO₂ pollution. The break values for population density are extracted from Baró et al. (2016), and the NO₂ concentration breaks consider the National Ambient Air Quality Standards (U.S. EPA 2014).

TABLE 8. Cross-tabulation matrix used to determine the demand for the ecosystem service “Air purification (O₃)” at the census block level.

Population density (inhabitants/ha)	O ₃ concentration (ppb)					
	<10	10–20	20–30	30–40	40–70	>70
<5	0	0	0	0	0	0
5–50	0	0.2	0.2	0.4	0.4	0.6
50–100	0	0.2	0.4	0.4	0.6	0.8
100–200	0	0.4	0.4	0.6	0.6	0.8
200–400	0	0.4	0.6	0.6	0.8	0.8
>400	0	0.6	0.8	0.8	0.8	1

Notes: Population density and mean O₃ concentration during the summer are used as exposure and hazard indicators to define the need for mitigating the risks associated with O₃ pollution. The break values for population density are extracted from Baró et al. (2016), and the O₃ concentration breaks consider the National Ambient Air Quality Standards (U.S. EPA 2014).

TABLE 9. Summary of the supply indicators used for each ecosystem service.

Ecosystem service	Supply indicator	Reference
Local temperature regulation	“Local climate indicator”—ratio between the local land surface temperature and the mean surface temperature of the green areas in the city	Schwarz et al. (2011)
Runoff mitigation	Water infiltration coefficient based on the curve number method	Cronshey (1986)
Air purification (NO ₂ and O ₃)	g·m ⁻² ·yr ⁻¹ absorbed by vegetation according to literature	Yang et al. (2008)

race and mismatch, we initially analyzed the distribution of people of color. In the context of this paper, we define people of color as those residents that are included in

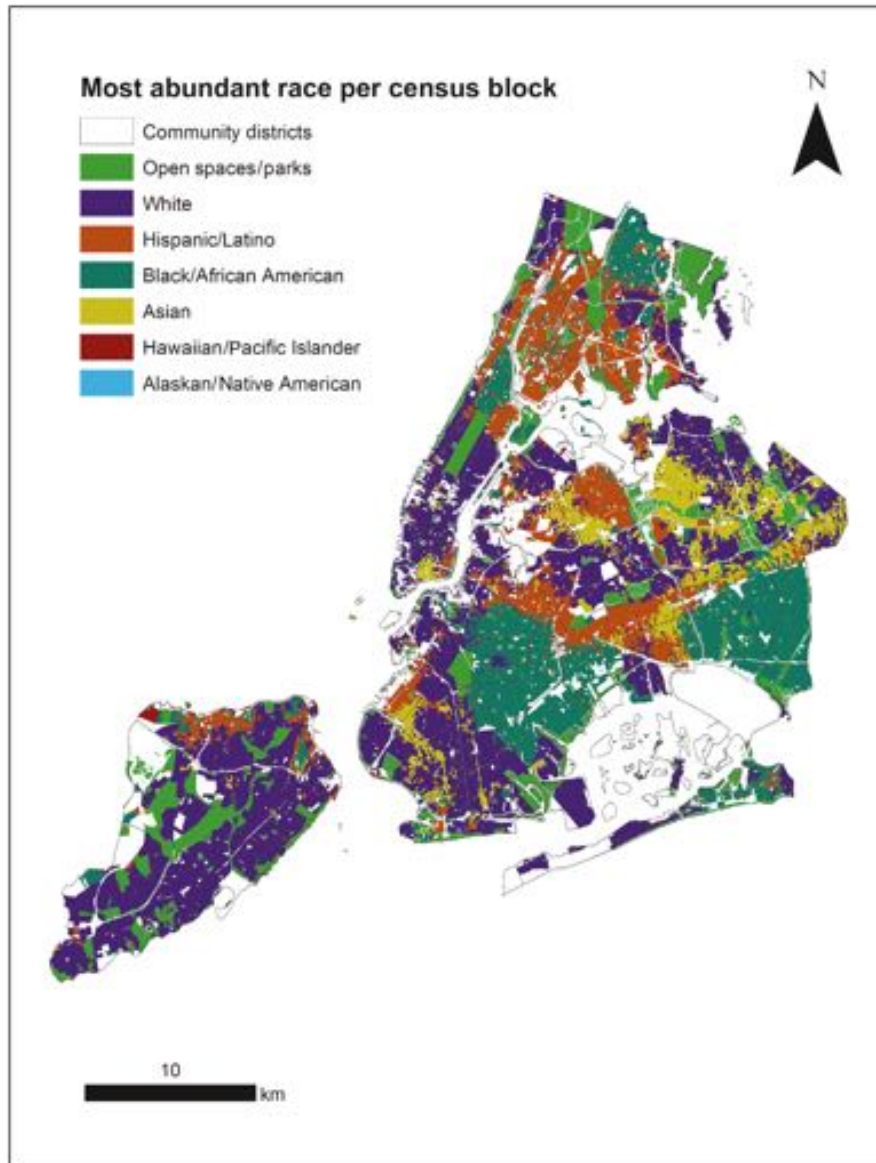


FIG. 4. Map showing the dominant race (race with highest percentage) per census block according to the 2010 decennial census (U.S. Census Bureau n.d. *a, b*).

the U.S. Census categories Hispanic/Latino, Black/African American, Asian, Alaskan/Native American, and Hawaiian/Pacific Islander. From here we then proceeded to analyze the distribution of the three most represented minority racial groups (Black/African American, Hispanic/Latino, and Asian) separately. Data on races and ethnicities per census block was retrieved from the Decennial Census 2010 through the data set titled “P9—Hispanic or Latino, and not Hispanic or Latino by race” (U.S. Census Bureau n.d. *a, b*) (Fig. 4). This data set presents a breakdown of the population per census block per ethnicity (Hispanic and not Hispanic) and race (White, Black/African American, Asian, Native American/Alaskan, and Hawaiian/Pacific Islander).

Data were incorporated into the geometries of the city’s census blocks obtained from the TIGER/LINE database via the data set “Special Release—Census Blocks with Population and Housing Unit Counts” in (U.S. Census Bureau n.d. *a, b*).

Median normalized income was retrieved from the American Community Survey (ACS) 5-yr estimates data set (2013–2017) (U.S. Census Bureau n.d. *a, b*). We relied on ACS for income data because it was not assessed in the 2010 decennial census (U.S. Census Bureau 2018). Since 2010, the ACS and the decennial census serve different purposes. The decennial census aims to generate an accurate population count that also captures age, sex, and race/ethnicity. The ACS, on the

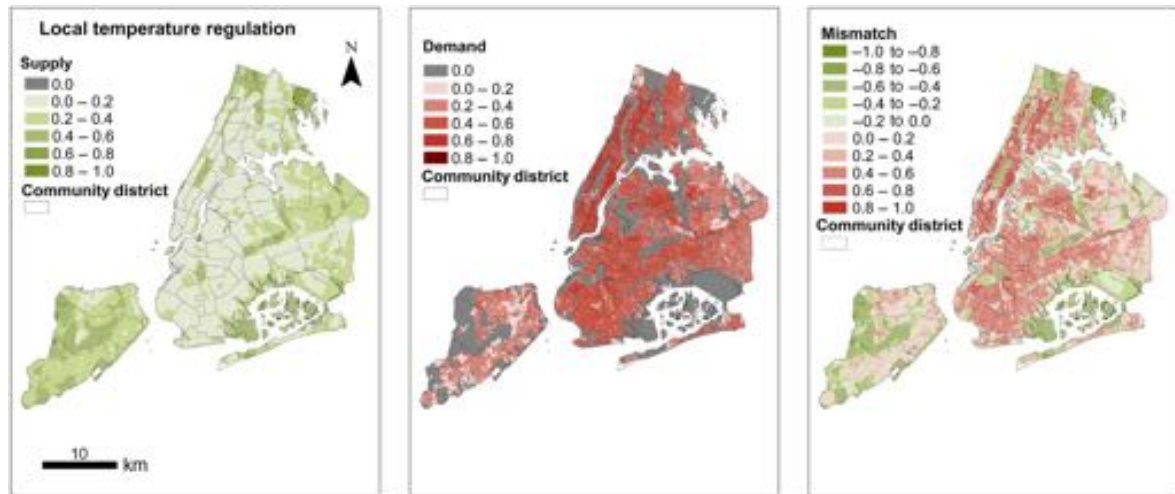


Fig. 5. Supply, demand, and mismatch maps at the census block level for the ecosystem service (ES) local temperature regulation. A composite figure; the rest of the ecosystem services assessed can be found in Appendix S1: Fig. S2.

other hand, focuses on estimating socio-demographic indicators based on a sample population that is smaller than that of the decennial census. While 100% of households are supposed to receive the decennial census form, only one in six households are sampled on a yearly basis to capture socio-demographic data. We used both data sets because, although income data does not exist in the decennial census 2010, it provides race/ethnicity data with higher accuracy. This data set was collected at a census block group level and was disaggregated into census block level through a spatial join, assuming that all the blocks within each group have the same median income.

To compare income and racial distribution across supply–demand mismatches, census blocks were spatially grouped by performing a hotspots analysis. Hotspots analysis is a widely adopted method in ecosystem services studies (Karimi et al. 2015, Morelli et al. 2017, Lorilla et al. 2019, Wang et al. 2019, Zen et al. 2019) because of its capacity to identify neighboring features that constitute an area with similarly outlying values. These areas may be targeted by different specific policies if they behave as hotspots (extremely high values) or coldspots (extremely low). The hotspots analysis was carried out in ArcMap 10.1 using the “Getis-Ord G_i^* ” tool (Environmental Systems Research Institute (ESRI) *n.d.*) to assess the mismatch value per ecosystem service. This procedure groups census blocks into spatial clusters based on their deviation from the values. Five cluster classes were identified (C1–C5, translating, respectively, into clusters with very significantly low, significantly low, nonsignificant, significantly high, and very significantly high mismatch values). The mean sociodemographic attributes per cluster class were then compared using a series of ANOVA analyses in R (version 3.6.1) (R Core Team 2020).

RESULTS

Spatial mismatch in ES supply and demand

Results show wide spatial variation in the supply, demand, and supply–demand mismatch per U.S. Census block (Fig. 5). ES supply across NYC varies according to the distribution of green space, with supply being highest in Staten Island, northwest Bronx, southern Brooklyn and eastern and southeastern Queens. For example, for local temperature regulation and air purification, the census blocks located between Central Park and Riverside Park and those that surround Prospect Park show higher supply values. ES demand is highest in most of Manhattan, central Bronx, central Queens, central Brooklyn, and the neighborhood of Greenpoint. Demand for local temperature regulation and runoff mitigation appear to have a similar distribution, likely due to the expected correlation between the hazard factors used in both services (Yuan and Bauer 2007).

Average values of supply, demand and supply–demand mismatch were calculated for each of the city’s CDs to compare the quantitative distribution of ES across NYC (Appendix S1: Fig. S3). A data set with the demand, supply, and mismatch values per CD is available in the Supporting Information file DataS1: Supply_Demand_Mismatch_Community_Districts. Even though supply, demand, and mismatch values of each ES follow similar paths, their trajectories are not identical, and have different mean values per CD. For instance, the supply for ES air purification has a higher value than for other ES, such as runoff mitigation. The summation of all the ES assessed is represented in Fig. 6, including neighborhood-scale examples.

Although supply for each of the ES assessed reaches its highest values in the CDs that fully overlap with major parks and natural areas (such as Central Park,

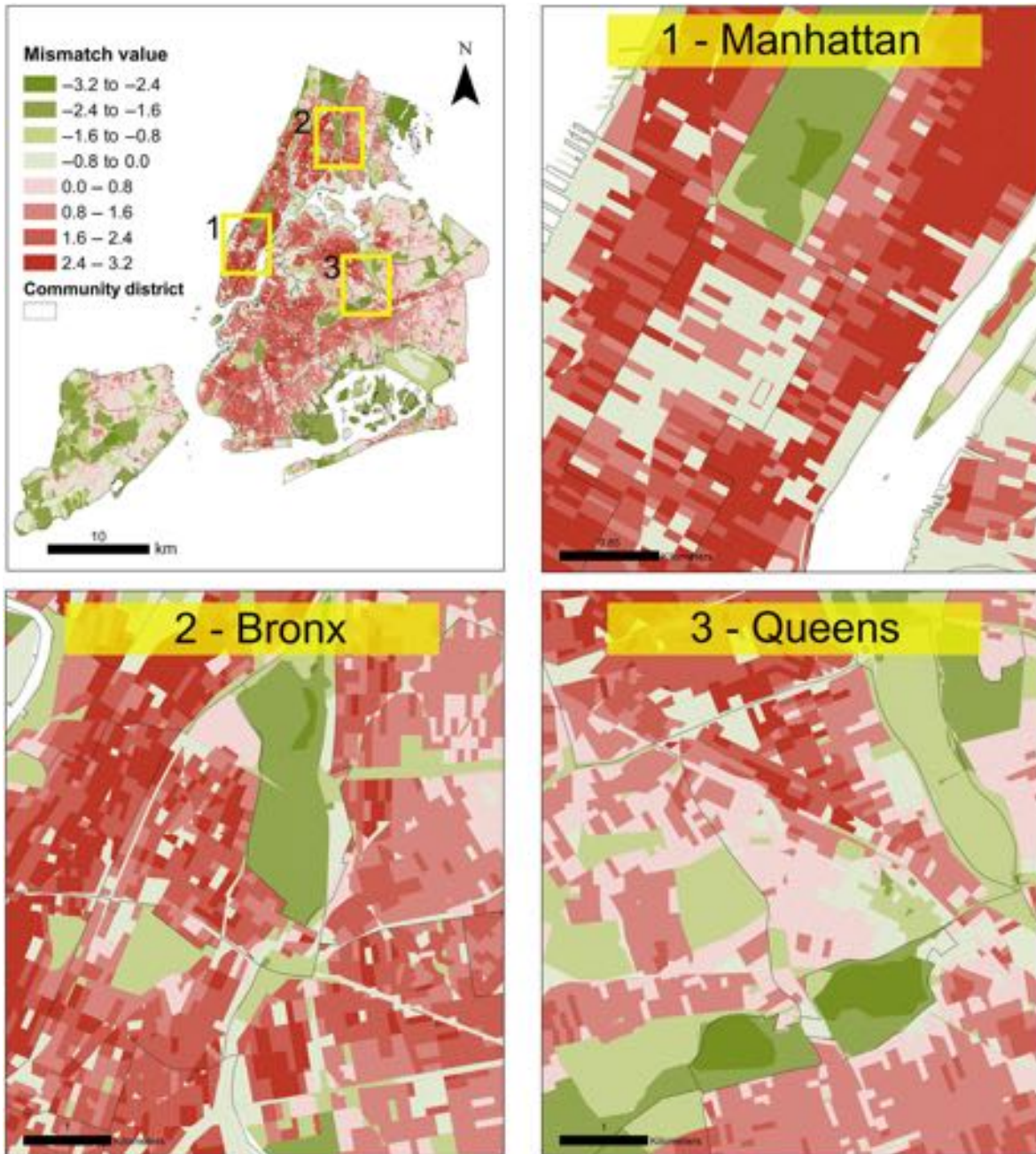


FIG. 6. Accumulated mismatch value of the four ecosystem services assessed and close ups. Important spatial nuances can be differentiated. For example, Midtown Manhattan (1) shows relatively low mismatch values because of the low population of census blocks occupied by office buildings, whereas the Upper East and West Side show high mismatch values, despite their wealthy population. In Bronx (2), a clear gradient by which central Bronx presents higher values than the edges of the borough. In Queens (3), the influence of parks in reducing the mismatch value of nearby blocks is visible in blocks like those situated in the center of the image.

Prospect Park, Pelham Bay Park and Forest Park), as well as the CDs located in Staten Island, demand shows opposite values. Given the absence of population, parks that showed high supply values show low or null demand. This is expected, of course, because low

population density drives demand, meaning lower demand in Staten Island as well as in CD 105-Midtown (Manhattan), where the prevalence of office buildings reduces the number of residents in this area. Due to its high supply values and low demand, the CDs located in

TABLE 10. Mean population density and hazard factor per demand value for each ecosystem service (ES).

	Demand for ES					
	0	0.2	0.4	0.6	0.8	1
Local temperature regulation						
Population density (inhabitants/ha)	0.2	29.7	71.0	203.6	618.3	
Mean temperature (°F)	85.3	82.6	86.5	86.7	85.3	
Runoff mitigation						
Population density (hab/ha)	1.5	35.54	89.1	213.7	543.7	592.6
Mean percentage impervious	30.3	25.8	37.4	47.2	48.2	84.3
Air purification (NO ₂)						
Population density (hab/ha)	0.5	41.9	105.3	251.5	604.7	
Mean NO ₂ (ppb)	22.8	17.8	22.0	25.6	28.2	
Air purification (O ₃)						
Population density (hab/ha)	0.7	23.9	61.0	199.4	618.77	
Mean O ₃ (ppb)	31.9	25.9	33.5	32.7	29.8	

Note: A demand of 1 was reached only in the service runoff mitigation. hab, inhabitants.

Staten Island show the lowest mismatch values among the inhabited CDs of the city.

Population density shows a consistent influence in the increase of demand for each ES, whereas the effect of the ES specific hazard factors on demand varies. As we show in Table 10, the hazard factor increases consistently across demand values only for the ES runoff mitigation and air purification—NO₂. For the ES local temperature regulation and air purification—O₃, the mean temperature per census block and the ozone concentrations are lower in the census blocks with higher demand.

Distributional environmental justice of ecosystem services in NYC

Spatial clusters of mismatch hotspots and coldspots are shown in Fig. 7. The differences in racial composition between mismatch clusters are significant and show that the average percentage of people of color is higher in the hotspots than in the coldspots (Fig. 8). On the other hand, the mean normalized income is lower in the mismatch hotspots. If we consider the three most represented races within people of color, Black/African American and Hispanic/Latino residents are increasingly present in higher mismatch clusters and their percentage is lowest in C1 (areas with the lowest mismatch), whereas Asian residents show no clear trend (Fig. 9).

DISCUSSION

Key findings—links with underlying injustices in NYC

The combination of ES supply and demand mapping shows consistent patterns of distributional environmental injustice across communities with different racial and income characteristics. Mismatch coldspots (low mismatch outliers) are inhabited by people with higher

incomes and characterized by lower percentages of people of color. As mismatch clusters shift from lower to higher supply–demand mismatch values, the proportion of people of color increases, and median income decreases. Hispanic/Latino inhabitants showed the most explicit trends of living in mismatch hotspots (high ES demand, low ES supply), with the proportion of Black/African American residents also being higher in mismatch hotspots. These results demonstrate that communities of color in NYC face a distributional injustice through lack of similar levels of access to the benefits provided by UGI in NYC when compared to predominantly white areas of the city.

The results of our analysis corroborate those obtained in other studies that examine relationships between environmental hazards, urban greenery, and socio-demographic variables such as race and income in American cities. For example, Hoffman et al. (2020) observed a consistent pattern in 108 formerly redlined cities where historically segregated neighborhoods showed higher surface temperatures and a less abundant tree canopy. Although the practice of redlining was banned in the late 1960s, the racial and economic differences between redlined and nonredlined neighborhoods remain visible (Jones 2017, Mitchell and Franco 2018) and have been maintained through zoning, public housing allocation, and subsidies distribution (Rothstein 2017), as well as public disinvestment processes (Stein 2019). Studies focused on green spaces and parks also have shown a consistently unequal distribution that especially affects Hispanic populations (Miyake et al. 2010, Rigolon 2016, Rigolon et al. 2018) and found that the presence of urban vegetation is strongly correlated with income and education (Nesbitt et al. 2019).

Here, our supply–demand mismatch approach also brings environmental justice considerations into a single methodological approach that combines measures of access to ecological benefits and exposure to

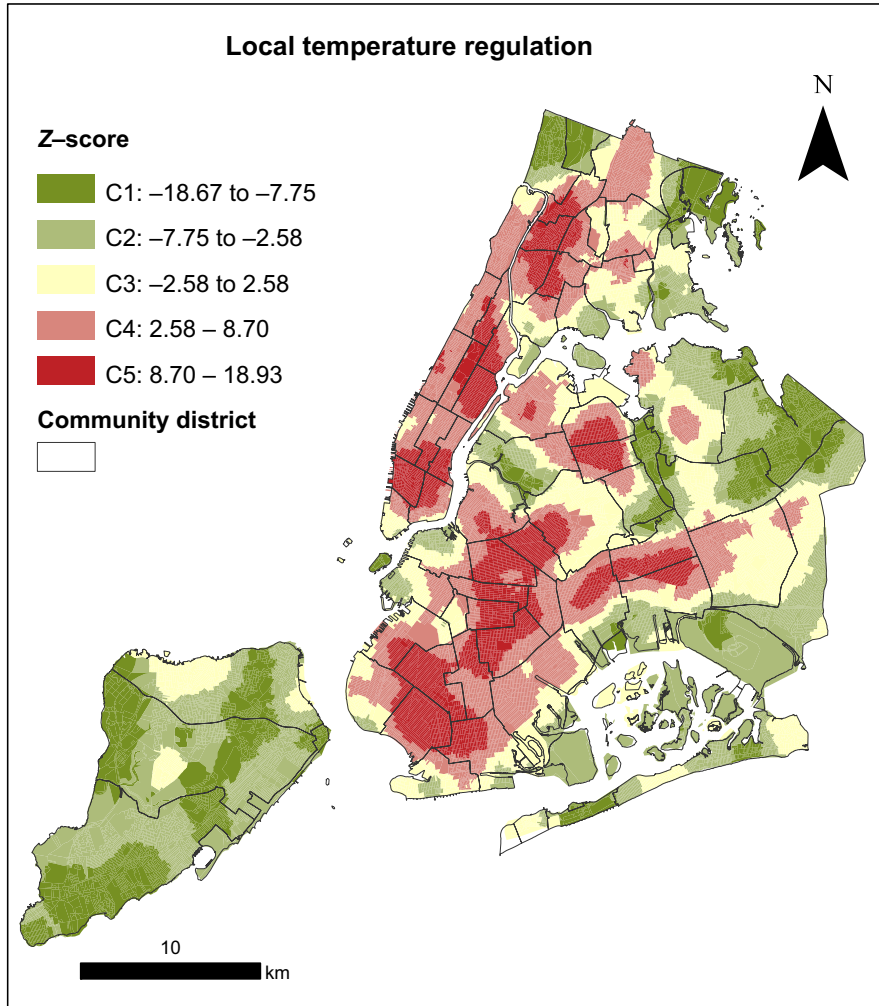


FIG. 7. Census blocks classified into hotspots and coldspots according to the Z-score obtained in the cluster analysis for the ecosystem service (ES) local temperature regulation. Clusters range from C1 (extreme lows, or coldspots) to C5 (extreme highs, or hotspots). Even though break values between different categories were set using a Jenks distribution, values 2.58 and -2.58 were set manually in order to keep a minimum degree of significance ($P < 0.01$). C3 corresponds to those census blocks that obtained a Z-score between 2.58 and -2.58, meaning that their P value is >0.01 . A composite figure with the rest of the ES assessed can be found in Appendix S1: Fig. S4.

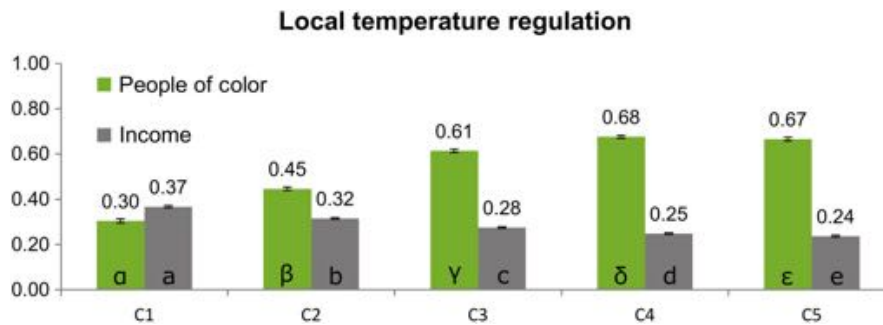


FIG. 8. Average proportion of people of color over the total population and relative income per mismatch cluster for the service local temperature regulation. Clusters range from C1 (extreme lows, or coldspots) to C5 (extreme highs, or hotspots). C3 refers to census blocks that do not belong to a high or low cluster based on statistical significance at $P > 0.01$. Latin and Greek letters indicate statistical significance across the clusters as per the ANOVA tests carried out. All the statistical significance tests returned P values below 0.001. A composite figure with the rest of the ecosystem services assessed can be found in Appendix S1: Fig. S5.

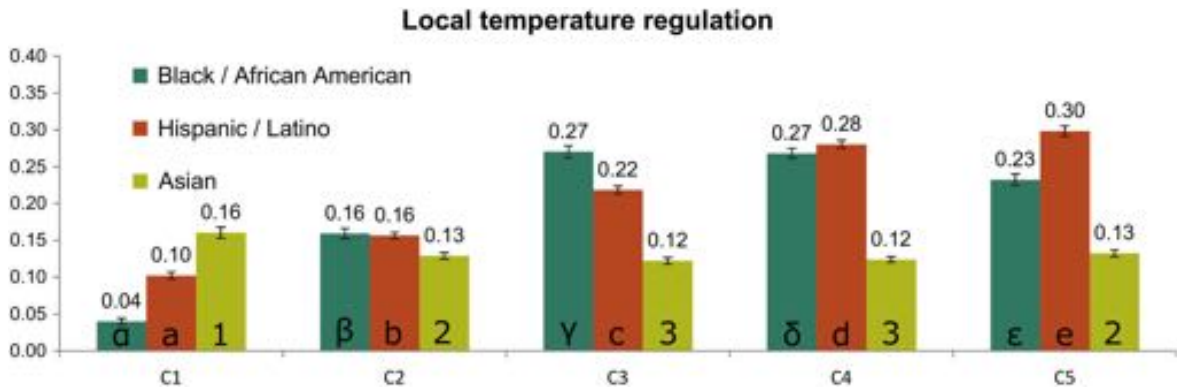


Fig. 9. Average proportion of disaggregated people of color, per mismatch cluster for the service local temperature regulation. Latin and Greek letters and numbers indicate statistical significance across the clusters as per the ANOVA tests carried out. All the statistical significance tests returned P values below 0.001, except for the proportion of residents being Asian when comparing C2–C3, C2–C4, C4–C5, and C4–C3 ($P < 0.01$). A composite figure with the rest of the ecosystem services assessed can be found in Appendix S1: Fig. S6.

environmental hazards. By relying on ES supply assessments instead of simply the distribution or quantity of urban green spaces, this study adds a new layer of ES complexity by considering the spatial variability of ecosystems' capacity to deliver benefits based on their ecological functions. Additionally, the mapping of supply–demand mismatch incorporates the idea that social need is not constant across space. Hence, mismatch mapping provides deeper insights about the areas in highest need of intervention and investment than a supply-only mapping exercise (Keeler et al. 2019a). We suggest that this approach could be a powerful planning tool for shifting NYC's current greening and hazard mitigation policies (New York City Department of Environmental Protection 2010, New York City 2017, 2019) in ways that enhance the impact of UGI implementation for benefiting areas most in need, and in ways that do not reproduce past injustices.

Location-based prioritization for UGI development is not new in NYC, given the limitation in resources and the need to maximize the cost efficiency of investments. However, the inclusion of justice or equity in the criteria used to site interventions varies across the city's plans. For example, the NYC Cool Neighborhoods Program (New York City 2017) specifies that new street tree investments need to be prioritized in areas that are not only hotter, but also that show higher social vulnerability to heat. Other programs, however, fail to incorporate explicit notions of environmental justice in their prioritization criteria. For instance, a recently passed tax abatement bill for green roofs establishes that buildings located in priority CDs may be subject to enhanced tax abatement (New York State Senate 2019). In this case, priority CDs are defined in relation to combined sewer overflow sewersheds and lack of green space, but do not use socio-demographic indicators to identify areas in higher need. The city's green infrastructure plan acknowledges the need to focus on “environmental

justice communities that need the additional public health and other sustainability benefits of green infrastructure” (New York City Department of Environmental Protection 2010:17), but does not specify how those communities may be identified. With this study, we hope that greening and resilience programs like these can benefit from a framework capable of highlighting the areas where ES mismatch is highest and describing their socio-demographic characteristics. With such insights, greening policies should be able to develop interventions that address the distributional injustices of UGI benefits in NYC while mitigating environmental hazards.

Research limitations and future iterations

Our approach for mapping ES demand was developed in accordance with the conceptualization of demand as “need for risk reduction” suggested by Wolff et al. (2015) and the cross-tabulation matrix developed by Baró et al. (2016). We used a traditional environmental justice lens, where risk is dependent on exposure to a hazard. The distribution of risk was then compared to the socio-demographic characteristics of differently exposed populations in order to assess the distributional justice of ES. This methodology is promising in that it is relatively simple to reproduce, and able to incorporate other social vulnerability indicators such as those considered by the Social Vulnerability Index of the Center for Disease Control and Prevention (Flanagan et al. 2018).

In addition, the results from mapping ES demand are driven by the hazard indicators chosen, the data used to quantify them, and the break values used in the cross-tabulation matrices. Mean temperature and concentrations of O_3 and NO_2 in the census blocks with the highest demand for each service are low compared to the highest break values considered in the cross-tabulation matrix. For example, the mean temperature in the census blocks with highly relevant demand for local

temperature regulation is 85.3°F, and the maximum temperature considered in the cross-tabulation matrix was >100°F. With such a low hazard factor, the only way census blocks can reach a demand value of 0.8 is with a higher population density. Thus, the interaction between hazards and population density drives demand index values. In future iterations, the thresholds defining ES demand breaks could be further developed in coordination with local authorities to consider the distribution of the values of the variables assessed. Further, we note that the supply, distribution, demand, and value of ES are influenced by other social and technological factors at the local scale (Andersson et al. 2015, Keeler et al. 2019b). The presence of technological or infrastructural services, as well as the configuration of the built environment, may influence the need for ES. For example, this study identified the major part of Manhattan as a mismatch hotspot, including the neighborhoods of the Upper West Side and the Upper East Side. These neighborhoods, which have predominantly wealthy, white populations, are characterized by a densely built environment with high-rise residential buildings and a wider adoption of residential air conditioning units than the rest of the city (Klein Rosenthal et al. 2014, Ito et al. 2018). Although taller buildings can contribute to an enhanced urban heat island effect (Ortiz et al. 2018), these may also be less affected by the cooling effect of street trees, and the widespread use of air conditioning may further reduce the need for ES in an area that was initially flagged as a mismatch hotspot. Hence, the effect of technological factors in the (supply and) demand for ES needs to be further considered in future iterations of this approach in order to ensure that the areas most in need for risk reduction are properly identified.

Regarding the assessment ES mismatch, it is important to point out the uncertainties that arise from comparing supply and demand in this study. To begin with, this approach maps mismatches by subtracting the normalized indicators for supply and demand, which are based on a combination of biophysical variables. Hence, this approach compares the relative values of supply and demand, or where demand is maximum and supply is minimum. Because this subtraction is carried out between directly noncomparable units, the resulting mismatch should not be understood as a scalar variable in which a mismatch of 0 means that 100% of the demand is being met by the supplied ES. Secondly, some of the risk indicators used to assess ES demand have an uncertain degree of influence by green areas. For example, the surface temperature data used already shows the impact of vegetation in lowering surface temperatures. This leads to an issue of double-counting when subtracting the supply of ES, because their effect is already accounted for. This uncertainty stems from the fact that, although ES supply is mapped using process-based models that represent biophysical processes, ES demand relies on actual measurements of temperature and air quality.

Although we follow other empirical studies that employ similar procedures to develop our ES demand and supply indicators (Burkhard et al. 2009, 2012, 2014, Kandziora et al. 2013, Burkhard and Maes 2017), the methods to calculate them need to be further integrated in order to ensure comparability. We propose two ways to ensure comparability in future studies. A first option is to rely on process-based modeling consistently for mapping both supply and demand. In this case, ES demand would be mapped as the populations that remain at risk after accounting for ES supply, which would be quantified by comparing the outcomes of simulating temperature, flooding, or air quality models with and without vegetation (see examples in Nowak et al. 2014, Glenis et al. 2018, Ortiz et al. 2018). A second, more complex, but widely needed option would be to combine or replace the modeling approaches used in ES supply mapping with actual measurements by relying on sensor networks that monitor the performance of UGI through time (Laney et al. 2015, Nitoslawski et al. 2019) in order to identify the actual performance of ES in moderating impacts of environmental hazards empirically.

Lack of data availability results in assumptions that unavoidably add uncertainty. For example, because of the lack of a continuous metric or a monitoring network, a constant relative humidity was assumed across the study area based on the three meteorological stations available. Regarding air purification, other relevant pollutants such as PM10 were not considered because of a lack of concentration data. The supply and demand for the ES local temperature regulation was assessed using surface temperature data because of the nonexistence of high-resolution local air temperature data, even though the differences between surface and air temperature are acknowledged (Bauer 2020). The lack of urban flooding observations in NYC means we, like others (Liquete et al. 2013), quantify ES demand for flood risk reduction using impervious surface data as a proxy indicator for flood risk.

It is important to recognize that indicators considered in this study to quantify ES supply, while widely used in ES literature (Haase et al. 2014), are simple representations of a complex reality that is affected by several factors at different scales. ES mapping approaches such as ours would benefit from incorporating the specific expertise of research fields that specialize in the underlying dynamics of each ES. For example, Eisenman et al. (2019) calls for ES scholars to incorporate epidemiological expertise when addressing the effect of urban ecosystems on air quality, which are known to be extremely complex and can also cause disservices through allergies or halogenated volatile organic compounds.

Environmental justice of this approach

There are important aspects that need to be considered in order to ensure that the approach presented in

this paper effectively contributes to the planning of UGI with an environmental justice lens. To begin with, interaction with local stakeholders, government officials, and community organizations is important to ensure that assessment of ES is relevant, and that the methods used are in accordance with the city's requirements. In addition, some local communities may prefer certain ES over others based on their own perceived needs or values (Wilkerson et al. 2018, Keeler et al. 2019b). Hence, a participatory process based on surveys, questionnaires, or participatory mapping would improve this research by defining weights with communities for each of the ES and environmental risks assessed. For instance, in Kremer et al. (2016), supply for several ES was aggregated into a single map using different weighting scenarios in a spatial multi-criteria analysis. In Depietri et al. (2018), a multihazard risk mapping in the city of NYC relied on local experts to develop a weighting criteria for different risk indicators. Incorporating participatory approaches, however, opens new questions regarding the procedural and recognitional justice (Walker 2009, Langemeyer and Connolly 2020) of UGI planning, including who should be consulted, who would be left out of the consultation, or how will participants be remunerated?

CONCLUSION

We provide a comprehensive, citywide and high-resolution understanding of the environmental justice implications of ES currently provided by UGI in NYC in relation to climate change adaptation and resilience priorities in the city. ES supply was mapped through a process-based model that quantifies supply based on a series of ecological proxies, while demand was framed as “need for risk reduction” and relied on social and physical factors. Results show that areas with a lower supply–demand mismatch tend to be populated by a larger proportion of white residents with higher median incomes, and areas with higher mismatch values, where need is high and supply is low, are present a population with lower incomes and a higher proportion of people of color. Analyses reveal clear examples of distributional environmental injustice in access to the climate-regulating benefits of ecosystem services provided by UGI in the city. Without improved analysis of current mismatches in supply and demand for critical climate regulatory ES, greening investments may exacerbate or even replicate historical and current environmental injustices and inequalities in American cities. Given the magnitude of the investments being made in NYC, but paralleled in many other cities globally, UGI development for climate change adaptation through ES delivery may be a critical opportunity to reduce the environmental justice burden on low-income and minority communities. We suggest that similar studies should be conducted in other cities and urban policy and planning should ensure that investments in such “nature-based”

solutions for climate change adaptation do not reinforce or exacerbate potentially existing environmental injustices.

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SUPPORTING INFORMATION

Additional supporting information may be found online at: <http://onlinelibrary.wiley.com/doi/10.1002/eap.2390/full>

OPEN RESEARCH

The data used as input for the manuscript are publicly available across different institutional websites. NYC Open Data was queried to obtain data sets with landcover data, <https://data.cityofnewyork.us/Environment/Landcover-Raster-Data-2010-3ft-Resolution/9auy-76zt> air pollutants concentration, <https://data.cityofnewyork.us/Environment/NYCCAS-Air-Pollution-Rasters/q68s-8qyv> and street centerlines. <https://data.cityofnewyork.us/City-Government/ NYC-Street-Centerline-CSCL-/exjm-f27b> Land surface temperature imagery was retrieved through USGS Earth Explorer using the criteria for date and cloud cover specified within our *Methods: Mapping ES demand* section. Census block geometries were retrieved from the Special Release from the TIGER/LINE source, <https://www.census.gov/geographies/mapping-files/time-series/geo/tiger-line-file.2010.html> and racial data were merged to the geometries using the table “P9—Hispanic or Latino, and not Hispanic or Latino by race,” which can be obtained through the Census Data platform through an advanced search by filtering the request for Race, 2010, census blocks data. <https://data.census.gov/cedsci/> Finally, data on income from the American Community Survey, as well as census block group geometries, were obtained through the TIGER/LINE FTP archive. https://www2.census.gov/geo/tiger/TIGER_DP/2017ACS/, State code = 36